1	Lower levels of human disturbance correspond with longer-term persistence of Endangered
2	Green Peafowl populations
3	
4	NITI SUKUMAL <sup>a,*</sup> , MATTHEW J. GRAINGER <sup>b</sup> and TOMMASO SAVINI <sup>a</sup>
5	
6	<sup>a</sup> Conservation Ecology Program, School of Bioresources&Technology, King Mongkut's University
7	of Technology Thonburi, 49 Soi Thian Thale 25, Bang Khun Thian Chai Thale Road, Tha Kham,
8	Bang Khun Thian, Bangkok 10150, Thailand.
9	<sup>b</sup> Norwegian Institute for Nature Research, P.O. Box 5685 Torgard, NO-7485 Trondheim, Norway.
10	*Author for correspondence; e-mail: niti_230@hotmail.com
11	
12	Summary
13	Galliformes are one of the most threatened groups of birds in Southeast Asia, with 27% of the
14	species threatened with extinction. Long term population viability and extinction probability
15	studies, at different levels of threat and management, are lacking due to weak life history data. This

study aimed to define the long-term viability and extinction risk of two populations of the

17 endangered Green Peafowl, facing different threat and protection levels, using Bayesian Population

18 Viability Analysis (BPVA), which requires less data than traditional methods. The results showed

19 an increasing trend in the Green Peafowl population in HuaiKhaKhaeng Wildlife Sanctuary

20 (western Thailand), with a high protection level and low disturbance, high probability of assuring

21 long-term persistence for the next 100 years; while the population in YokDon National Park

22 (southcentral Vietnam), with a high habitat disturbance level and hunting pressure, is predicted to

decline and has a high probability with 99% of extinction by 2097. Also, the BPVA showed

24 minimum viable population (MVP) sizes of 250 and 450 calling males are required for the

25 HuaiKhaKhaeng and YokDon populations respectively, both assuring high probability of long-term

26	persistence if the minimum numbers of males are available. However, the population size of 219
27	calling males at YokDon during the 2013 survey is lower than the MVP threshold of 450 calling
28	males, which suggests the species has a low probability of long-term persistence in the area. Despite
29	the limited life history data, BPVA predicted future population trend under site-specific conditions,
30	and the information could be useful for proper management and population restoration.
31	
32	Keywords: Galliformes, Minimum Viable Population, Bayesian Population Viability Analysis,
33	HuaiKhaKhaeng Wildlife Sanctuary, YokDon National Park
34	
35	Introduction
36	Vertebrate diversity has drastically declined over the past decades in tropical regions (Hoffmann et
37	al., 2010). Southeast Asia in particular has a high proportion of species at risk of extinction within
38	the next 50 years (Sodhi et al. 2004, Laurance 2006). This is due primarily to the degradation of
39	natural habitat (Sodhi et al. 2010) and overhunting (Harrison et al. 2016). Birds particularly at risk
40	with about 10% of the 2,696 species are designated as threatened with extinction, including about
41	27% of the 76 Galliformes species (IUCN 2016). However, there are no empirical data on the
42	extinction probability of the Galliformes species in Southeast Asia (Grainger et al. 2018).
43	To support conservation and management decisions, it is imperative to determine,
44	quantitatively, the extinction probability of threatened species in the region. In the past, this has
45	been achieved using Population Viability Analysis (PVA), which predicts long-term population
46	survival, extinction risk and associated threats within a specific environment (Boyce 1992, William
47	et al. 2002). PVA analyzes how population growth parameters, such as initial population size,
48	population growth rate, survival rate, productivity and predation rate are affected by environmental
49	factors, such as habitat quality, disturbance level and protection level (Ruggiero et al. 1994).

50 Understanding the variables affecting the survival trend of a population is crucial to defining the

actions needed for the species long term management (Brook et al. 2000). Besides providing
important details on the probability of extinction, PVA can be used to determine the Minimum
Viable Population (MVP), defined as the minimum number of individuals of a species inhabiting a
given area needed to guarantee survival of the population over a specific time frame (Shaffer 1981);
higher extinction rate is typically observed in small populations in comparison with large ones
(Schoener and Spiller 1987). In addition, MVP can be used to estimate the minimum area required
for any given carrying capacity.

Although PVAs have potential to provide critical information for conservation planning, 58 some studies have questioned the reliability of PVA due to large uncertainty in model outcomes 59 (Flather et al. 2011). Viability is generally predicted using population models such as those 60 incorporated in the Vortex software (Lindenmayer et al. 1995), that requires multiple parameters 61 62 (e.g. productivity data, population growth rate and survival rate), however some parameters are not often available, as they require long-term studies (Boyce 1992). Gilpin and Soule (1986) suggested 63 that a study of 12 generations, at least, is required for a robust extinction risk prediction. 64 Unfortunately, most ecological studies are carried out over a three-year period (Tilman 1989, 65 Baskin 1997), providing data that often underestimate the extinction risk (Reed et al. 2003). 66 Generally, the life history of a species is dependent on environmental factors and management 67 strategies (Flather et al. 2011) and therefore, each population within an area ought to be separately 68 considered while performing a PVA analysis. Lack of relevant data has been reported as the main 69 hindrance to viability analysis for Galliformes (Fuller and Garson 2000) and to date, only Bro et al. 70 (2000) and Zhang and Zheng (2007) have conducted PVA studies for Galliformes in Southeast 71 72 Asia. However, by running PVA in a Bayesian framework, we can develop predictive models using 73 information on observed population abundance, for a period of at least two years, and simulate population information for a given period by random sampling within the range of the observed 74 75 population abundance data from two observed periods. This approach allows the estimation of a

credible range of future population abundance and predicts the probability of a future trendemploying only monitoring abundance data.

The Endangered Green Peafowl (Pavo muticus) is among the most threatened species in 78 79 Southeast Asia due to high hunting pressure and habitat destruction (McGowan et al. 1999). The original species distribution covered the whole dry-forest areas (e.g. deciduous, dry dipterocarp, 80 pine forest) in northeast India, Bangladesh, Yunnan (southwest China), Myanmar, Thailand, Laos, 81 Vietnam, Cambodia, peninsular Malaysia and Java, Indonesia (McGowan et al. 1999). However, 82 recent modeling shows a severe decline in range currently scattered into small patches highlighted 83 in six population strongholds, including southern Myanmar; northern, northwestern and western 84 Thailand; and northern and eastern Cambodia (Sukumal et al. in press). Recently several population 85 estimates have been made for some of the strongholds (Sukumal et al. 2015, Nuttall et al. 2016, 86 Loveridge et al. 2017, Sukumal et al. 2017). The loud mating call by males makes detection of 87 Green Peafowl relatively easy compared to other Galliformes species in the region (Suwanrat et al. 88 2015), making them a good species to investigate the predictive response of ground dwelling birds, 89 90 especially in dry forest habitat, to different levels of threat and management strategies.

91 Therefore, the aims of this work were to estimate the long term viability and extinction risk of Green peafowl using only information on the observed population abundance in two years 92 93 separated by a long interval. We use a Bayesian model to provide estimates of uncertainty as the data we have are limited (the estimates of uncertainty would be narrower in a frequentist model). In 94 addition, we do not have the requisite ecological data (vital rates such as survival, fecundity, and net 95 immigration) to build a more complex model or use stand-alone PVA software (e.g. Vortex). We 96 97 use two populations of Green Peafowl facing different levels of threat as a case study. Specifically, 98 we 1) estimated the population growth rate at two sites, showing different protection level, where two surveys with count data were conducted at 23 year intervals at a first site and 15 year intervals 99 at a second site (Sukumal et al. 2015, Sukumal et al. 2018) and we projected the results into the 100

future to see for how long the populations will persist in protected areas facing different threats and
2) defined the minimum viable population (MVP) size of Green Peafowl required for a long-term
persistence of the species in a given area.

104

### 105 Study area

106 This study was focused on two sites with different levels of protection and disturbance to the Green Peafowl population. The first site was in HuaiKhaKhaeng Wildlife Sanctuary (HKK), western 107 Thailand (15°36' N 99°19' E), with increased patrolling and reduced hunting pressure that led to an 108 increase in the population of Green Peafowl between 1992 and 2015 (Sukumal et al. 2017). The 109 HKK is connected to ThungYaiNaresuan Wildlife Sanctuary and this Western Forest Complex is 110 designated as a UNESCO World Heritage Site. It covers a total area of 2,780 km<sup>2</sup> with an 111 altitudinal range of 200 to 1,600 m, an annual temperature range of 8 to 38°C and a mean annual 112 rainfall of 1,375 mm (Simcharoen et al. 2014). The dry season in HKK occurs between November 113 and April, with a mean rainfall of 298 mm; and the wet season occurs between May and October, 114 with a mean rainfall of 1,088 mm. The sanctuary consists of mixed deciduous forest (48%), dry 115 evergreen forest (25%), hill evergreen forest (13%) and dry dipterocarp forest (7%) (Simcharoen et 116 al. 2014). The information on population estimation was derived from two independent surveys 117 during a 23-year period, comprising a historical survey between 1989 and 1992 (Simcharoen et al. 118 1995, McGowan et al. 1999) and a follow-up survey in 2015 (Sukumal et al. 2017). 119 The second site was in YokDon National Park (YDN), southcentral Vietnam (12°47′ - 13° 120 00'N, 107'29' - 107'50'E), with a high level of habitat disturbance and hunting pressure that led to a 121 drastic decline in the Green Peafowl population between 1998 and 2013 (Sukumal et al. 2015). 122 123 The YDN is located in DakLak Province, which borders Cambodia and has a total area of 1,155 km<sup>2</sup>, mainly dominated by dry deciduous forest (75% of total area). The area is relatively flat with 124 125 an elevation of about 200m and high spots of up to 474m (Sterling et al. 2006). This area comprises

the majority of suitable habitat for Vietnam's remaining Green Peafowl population (Brickle 2002).

127 However, a recent survey in 2013 reported a much reduced Green Peafowl population in YDN

128 (Sukumal et al. 2015), compared to the population number in 1998 (Brickle et al. 1998).

129

#### 130 Methods

131 *Population monitoring data* 

The estimation of the Green Peafowl population number in HKK was based on two historical 132 datasets on recorded calling males, comprising 256 (Simcharoen et al. 1995) and 225-300 calling 133 males (McGowan et al. 1999); and the dataset from a follow-up survey, comprising 434 calling 134 males with a range of 373 to 512 (95% Confidence Interval estimated using DISTANCE) (Sukumal 135 et al. 2017). These datasets were used to model the population abundance over the last hundred 136 years. For YDN, a historical dataset comprising an estimated 475 calling males, ranging from 391 137 to 559 (Brickle et al. 1998) and an estimated 219 calling males, ranging between 101 and 479 138 calling males, from a recent survey (Sukumal et al. 2015) were used as input for the model. 139

140

## 141 *BPVA framework*

The analysis was conducted using a Bayesian state-space model. We analyzed the population 142 datasets for HKK with a high level of protection and low level of disturbance (Sukumal et al. 2017), 143 and YDN with an observed high disturbance level (Sukumal et al. 2015) to illustrate the potential 144 differences in extinction probability. We chose a state-space approach to account for both sampling 145 and demographic uncertainty. Ideally, one would incorporate monitoring data over a longer time 146 period, but such data is unavailable for Green Peafowl. Hence, for the simulation procedure, a 147 148 constant rate of growth was assumed throughout the projected time series. The initial population size was inputted in range from 0 to the initial population estimate. We used estimated historical 149 150 population number at each site as a maximum initial population size in the analysis. The population

151 estimates from the historical and recent surveys were used to calculate annual population abundances in HKK, from 1992 to 2015, and in YDN, between 1998 and 2013. These annual 152 population abundances were used to estimate average population growth rates and the average 153 154 population growth rate values were then used to estimate the future population abundance trend over the next 100 years (1992-2092 for HKK, and 1998-2098 for YDN). We ran BPVA under the 155 assumption of exponential growth rate when no constraints were applied because the estimation of 156 carrying capacity (k) is highly uncertain, especially for a species that has been hunted 157 unsustainably, resulting in a sparse population distribution across the entire range. 158 159 Data analysis 160 Posterior distribution of parameters was conducted using Markov chain Monte Carlo (MCMC) 161

simulations in the program JAGS, implemented using the R (R Development Core Team 2014)

package "R2jags" (Su and Masanao 2015). The analysis was run using three independent chains for
50,000 iterations after a burn-in of 10,000 iterations.

MVP was determined following BPVA approach (Saunders et al. 2018) whereby each population number was inputted in a loop procedure as the mean of recent population sizes. Next, the lowest population number that showed a stable or increase future population trend after 50 years was used to simulate the population abundance.

169

## 170 **Results**

171 Population viability at different disturbance levels

172 In HKK (western Thailand), with a high level of protection and low disturbance, the Green Peafowl

population size was stable over the study period, as evidenced by the BPVA estimated population

abundance of 209 (credible interval or CI predicted by model = 26-255) and 261 (CI = -49 - 411)

175 calling males in 1992 and 2015 respectively, with an overlapping CI. If the current protection and

disturbance levels in the area remains, the population size will increase to 1,516 calling males (CI =
896-2,081) by 2091 (Figure 1).

In contrast, the population size in YDN (southcentral Vietnam), with a high level of habitat disturbance and hunting pressure, drastically declined between 1998 and 2013. The estimated population abundance, along with the corresponding CI in 1998 and 2013 were 458 (CI = 410-474) and 111 (CI = 33-165) calling males, respectively. If the status quo remains going forward, without any conservation or management improvement, the species has a 63% chance of going extinct, based on our model, as evidenced by the estimated population size of 0.95 (CI = 0.17-3.93) in 2052 (Figure 2).

185

186 *Minimum Viable Population (MVP)* 

In HKK, the smallest population number that could maintain long-term population survival (i.e. MVP) for 76 years from last survey, i.e. from 2016 to 2091, or for the 100-year period between 189 1992 and 2091, is  $250 \pm 124$  (SD) calling males, assuming the current protection and disturbance 190 levels, comparable to conditions in 1992, are maintained going forward. However, this number is 191 lower than the estimated initial historical population size of 256 calling males (Simcharoen et al. 192 1995) (Figure 3).

In contrast, for YDN, the smallest initial population size that could maintain a long-term population survival (i.e. MVP) for 84 years from last survey (i.e. from 2014 to 2097) is  $450 \pm 145$ calling males (Figure 4). This number lies within the initial historical population range of 391 to 559 (Brickle et al. 1998) and is higher than the recent population estimate of 219 calling males, indicating that Green Peafowl population is currently at a level of population size that does not appear to be sustainable, given demographic influences in the region.

199

### 200 Discussion

201 We conducted our analysis employing only data on population abundance from two timeframes at two different sites, with different protection and disturbance levels, to simulate the historical, recent 202 and future (next 100 years) population trends. Results from the HuaiKhaKhaeng Wildlife 203 204 Sanctuary, which has a high level of protection and low disturbance, suggest a positive Green Peafowl population growth, assuring long-term population persistence for the next 70 years. An 205 206 MVP size of 250 calling males was determined to guarantee the species long-term persistence in the area. In contrast, the Green Peafowl population in YokDon National Park, suffering high habitat 207 disturbance and hunting pressure, drastically declined and our analysis predicts the population has a 208 63% chance of going extinct in the year 2052, in the absence of intervention measures. Our model 209 210 estimated that a population level of 450 calling males were the lowest level at which a stable population was predicted under the assumptions of our model, which included an assumption of 211 exponential growth. However, the latest survey, in 2013, showed the population number is lower 212 than this MVP, and we infer this to be strong evidence that the species may be at risk of local 213 extinction. 214

215

# 216 Using population abundance information with BPVA

Our ecological knowledge of Green peafowl is still limited in terms of life history parameters such 217 as annual productivity despite the availability of efficient methods that increase their detection 218 probability, compared to other forest Galliformes species (Suwanrat et al. 2015). In the absence of 219 the life history data, the state-space modelling approach allowed us to simulate the current and 220 future population abundance of the species using historical monitoring data. However, because we 221 lacked empirical productivity and survival information, our model's outputs were characterized by 222 223 high uncertainty (i.e. wide Credible Interval of posterior abundance). Therefore, we recommend the integration of life history information, when available, to improve the precision of the analysis. 224

Given the difficulty in obtaining life history data, our basic model is a promising tool forconservation planning.

Population viability analysis has been scarcely applied to study Galliformes species, 227 especially in Southeast Asia. For the whole Asia region, only one case study that collected the long-228 term life history data (from 1982 to 2004) of Cabot's Tragopan (Tragopan caboti) in Wuyanling, 229 south-east China, and VORTEX was used in the study to estimate population viability (Zhang and 230 Zheng 2007). VORTEX requires substantial life history information (e.g. population growth rate, 231 survival rate) for conducting PVA. Nevertheless, by integrating the limited data on population 232 abundance with Bayesian approach, reliable prediction of the future viability of the population, and 233 appropriate management actions could be recommended. 234

235

### 236 *Conservation and management implication*

As expected, the two protected areas show a marked difference in the disturbance and protection 237 levels. HuiKhaKhaeng has a high level of protection with minimum disturbances. The effective 238 patrolling system covering the entire area has resulted in a reduction of hunting pressure. Moreover 239 the sanctuary has reduced human activities due to the prohibition of settlement within the area. 240 The good management in HuaiKhaKhaeng Wildlife Sanctuary has reportedly led to an increase in 241 population of various endangered species, e.g. Tiger (*Panthera tigris*) (Duangchantrasiri et al. 242 2016), Banteng (Bos javanicus) and Green Peafowl (Sukumal et al. 2017). If the current 243 management system is maintained, the population levels of Green Peafowl have high probability of 244 being maintained for the next 70 years. However, monitoring of Green Peafowl population 245 abundance should be conducted every 5 years, which is the age of sexual maturity for males 246 247 (Madge and McGowan 2002). To assure long-term persistence of the Green Peafowl population in the site managers should aim to maintain at least 250 calling males. By contrast, YokDon is 248 249 characterized by high habitat disturbance and hunting pressure. Overgrazing by domestic cattle and

bushfire have led to the deterioration of vegetation cover, resulting in the loss of foraging and 250 nesting sites for Green Peafowl. These disturbances, in addition to a high hunting pressure, 251 contributed to the drastic decline in Green Peafowl population in the area (Sukumal et al. 2015). A 252 253 study conducted in 2013 revealed the counted population of Green Peafowl in YokDon was below the level predicted by the model as the one needed to sustain it in the long term following the 254 255 current growth rate, indicating that the population may be at risk of local extinction. Measures such as restriction of human activities and effective patrolling are crucial to improving the Green 256 Peafowl population. YokDon is strategically important for the survival of the species in Southeast 257 Asia as it is connected to Mundulkiri Protected Forest at eastern part of Cambodia, which is home 258 to a large population of Green Peafowl (Sukumal et al. in press). 259

Generally, native wildlife populations are facing different threat levels and management 260 261 strategies are urgently required to ensure their survival. Understanding how a population responds to site-specific conditions is key to setting up proper management plans for effective population 262 restoration. Our results clearly show that protection level affects long term survival of species, we 263 can see that with improved protection populations tend to increase, while with absence or reduce 264 protection population decrease. Therefore, projecting future protection level, when those data are 265 available, should be included when performing viability analysis. Information on population 266 abundance can be useful to investigate the population viability and predict plausible future 267 population trend. Considering the relative ease of detecting calling males during the breeding 268 season, Green Peafowl could be used as indicator species to investigate the response of ground-269 dwelling birds to human-induced threats in their dry forest habitat. Information on population 270 abundance could be integrated with BPVA to predict future population trends under site-specific 271 272 conditions. Moreover, we expect that this analysis can be applied to other species where only population abundance and not life history information is available. The BPVA can also be applied 273

274	to a broader scale for determining suitable management strategies especially for highly threatened
275	species.
276	
277	Acknowledgements
278	The authors acknowledge King Mongkut's University of Technology Thonburi for funding
279	Postdoctoral Fellowship to Niti Sukumal. Thanks to O. Nnaemeka for improving an English in this
280	manuscript.
281	
282	References
283	Baskin, Y. (1997) Center seeks synthesis to make ecology more useful. Science 275: 310–311.
284	
285	Boyce, M.S. (1992) Population Viability analysis. Annual Review in Ecology and Systematic 23:
286	481-506.
287	
288	Brickle, N.W., Cu, N., Quynh, H.Q., Cuong, N.T., and Van San, H. (1998) The status and
289	distribution of Green Peafowl Pavo muticus in Dak Lak province, Vietnam. BirdLife International
290	Vietnam Programme/Institute of Ecology and Biological Resources (Conservation Report 1),
291	Hanoi, Vietnam.
292	
293	Brickle, N.W. (2002) Habitat use, predicted distribution and conservation of Green Peafowl (Pavo
294	muticus) in Dak Lak Province, Vietnam. Biological Conservation 105: 189-197.
295	
296	Bro, E., Sarrazin, F., Clobert, J., and Reitz, F. (2000) Demography and the decline of the grey
297	partridge Perdix perdix in France. Journal of Applied Ecology 37: 432-448.
298	

- Brook, B.W., O'Grady, J.J., Chapman, A.P., Burgman, M.A., Akcakaya, H.R., and Frankham, R.
  (2000) Predictive accuracy of population viability analysis in conservation biology. *Nature* 404:
  287-385.
- 302
- 303 Flather, C.H., Hayward, G.D., Beissinger, S.R., and Stephens, P.A. (2011) Minimum viable
- populations: is there a 'magic number' for conservation practitioners? *Trends in Ecology and Evolution* 26: 307-316.

307 Fuller, R.A., and Garson, P. J. (2000) Pheasants. Status Survey and Conservation Action Plan

308 2000-2004. WPA/BirdLife/SSC Pheasant Specialist Group. IUCN, Gland. Switzerland and

309 Cambridge, UK and the World Pheasant Association, Reading, UK.

310

- Gilpin, M.E., and Soule, M. E. (1986) *Minimum viable population: processes of extinction*. Pp. 19–
- 312 34 in M. E. Soule', ed. Conservation biology: the science of scarcity and diversity. Sunderland,
- 313 MA: Sinauer.
- 314
- Grainger, M.J., Garson, P.J., Browne, S.J., McGowan, P.J.K., and Savini, T. (2018) Conservation
  status of Phasianidae in Southeast Asia. *Biological Conservation* 220: 60-66.

317

- Harrison, R.D., Sreekar, R., Brodie, J.F., Brook, S., Luskin, M., O'kelly, H., Rao, M., Scheffers, B.,
- and Velho N. (2016) Impacts of hunting on tropical forests in Southeast Asia. *Conservation Biology*30: 972-981.

321

Hoffmann, M. et al. (2010) The impact of conservation on the status of the world's vertebrates. *Science* 330: 1503–1509.

- 325 IUCN (2016) The IUCN Red List of Threatened Species. Version 2016-2. Available at:
- 326 http://www.iucnredlist.org. [accessed 31 October 2016].

327

Laurance, W.F. (2006) Have we overstated the tropical biodiversity crisis? *Trends in Ecology and Evolution* 22: 65–70.

330

Lindenmayer, D.B., Burgman, M.A., Akçakaya, H.R., Lacy, R.C., and Possingham, H.P. (1995) A
review of the generic computer programs ALEX, RAMAS/space and VORTEX for modelling the
viability of wildlife metapopulations. *Ecological Modelling* 82: 161–174.

334

- Loveridge, R., Kidney, D., Ty, S., Eames, J. C., Borchers, D., Kidney, D., and Borchers, D. (2017)
  First systematic survey of Green Peafowl Pavos muticus in north eastern Cambodia: reveals a
  population stronghold and preference for disappearing riverine forests. *Cambodian Journal of Natural History 2017*: 157-167.
- 339
- 340 Madge, S., and McGowan, P.J.K. (2002) *PHEASANTS, PARTRIDGES AND GROUSE*.
- 341 Christopher, Helm, London.

342

343	McGowan, P.J.K.	, Duckworth,	J.W., V	Wen, X.,	Van Balen,	В.,	Yang, X.	, Khan,	M.K.M.,	Yatim,
-----	-----------------	--------------	---------	----------	------------	-----	----------	---------	---------	--------

S.H., Thanga, L., Setiwan, I., and Kaul, R. (1999) A review of the status of the Green Peafowl *Pavo muticus* and recommendations for future action. *Bird Conservation International* 9: 331-348.

- Nuttall, M., Nut, M., Ung, V., and O'Kelly, H. (2016) The first abundance estimates for the
- endangered Green Peafowl Pavo muticus in Cambodia: identification of a globally important site

349 for conservation. *Bird Conservation International* 27: 127-139.

350

351	Reed, D. H., O'Grady, J.J., Brook, B.W., Ballou, J.D., and Frankham, R. (2003) Estimates of
352	minimum viable population sizes for vertebrates and factors influencing those estimates. Biological
353	Conservation 113: 23-34.

- 354
- Ruggiero, L., Hayward, G., and Squires, J. (1994) Viability analysis in biological evaluations:
  concepts of population viability analysis, biological population, and ecological scale. *Conservation Biology* 8: 364–372.
- 358

Saunders, S.P., Cuthbert, F., and Zipkin, E. F. (2018) Evaluating population viability and efficacy
of conservation management using intergrated population models. *Journal of Applied Ecology* 7:
1380-1392.

362

- Shaffer, M.L. (1981) Minimum population sizes for species conservation. *BioScience* 31: 131–134.
  367
- 368 Simcharoen, S., Thongnamchaima, B., Sukmasuong, R., Thobmongkol, P., Khoothong, R.,
- 369 Thobmongkol, P., Khoothong, M., Sunthran, C., Mheesangpraew, Y., Thongooppagarn, W., and
- 370 Singkram, P. (1995) Population and distribution range of green peafowl (*Pavo muticus*) in Huai
- 371 Kha Khaeng Wildlife Sanctuary. *Wildlife Journal of Thailand* 4: 43-48 (In Thai).
- 372
- 373 Simcharoen, A., Savini, T. Gale, G.A., Simcharoen, S., Duangchantrasiri, S., Pakpien, S., and

<sup>Schoener, T.W., and Spiller, D.A. (1987) High population persistence in a system with high
turnover.</sup> *Nature* 330: 474–477.

Smith, J.L.D. (2014) Female tiger Panthera tigris home range size and prey abundance: important
matrix for management. *Oryx* 48: 370-377.

376

- Sodhi, N.S., Koh, L.P., Brook, B.W., and Ng, P.K.L. (2004) Southeast Asian biodiversity: an
  impending disaster. *Trends in Ecology and Evolution* 19: 654–660.
- 379
- Sodhi, N.S., Posa, M.R.C., Ming Lee, T., Bickford, D., Koh, L.P., and Brook, B.W. (2010) The
  state and conservation of South-east Asian biodiversity. *Biodiversity and Conservation* 19: 317–
  328.

383

Sterling, E.J., Hurley, M.M., and Le Duc, M. (2006) *Vietnam: A Natural History*. Yale University
Press, New Haven, London.

386

Su, Y., and Yajima, M. (2015) *R2jags: Using R to Run 'JAGS*'. Rpackage version 0.5-7. URL: *http://CRAN.R-project.org/package=R2jags*.

389

- 390 Sukumal, N., McGowan, P.J.K., and Savini, T. (2015) Change in status of green peafowl Pavo
- 391 *muticus* (Family Phasianidae) in south Central Vietnam: A comparison over 15 years. *Global*
- *Ecology and Conservation* 3: 11-19.
- 393

394 Sukumal, N., Dowell, S D., and Savini, T. (2017) Micro-habitat selection and population recovery

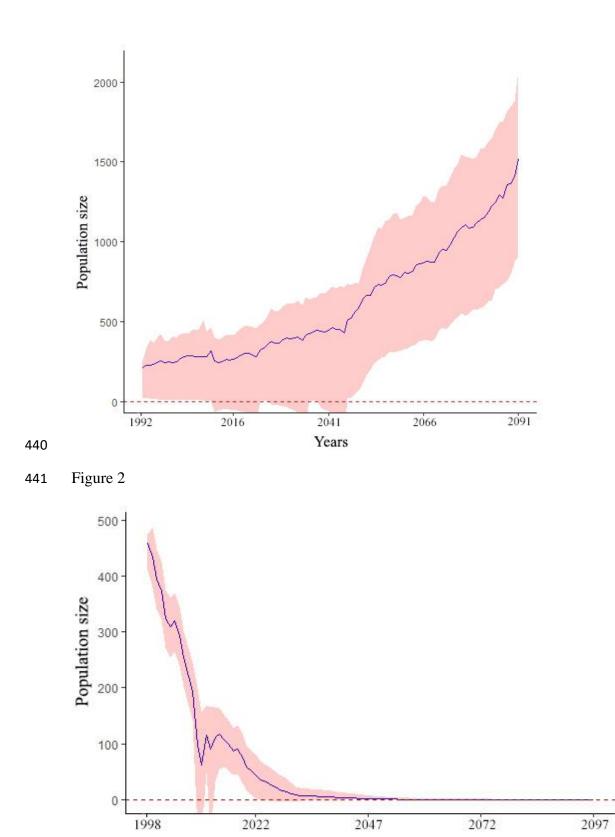
- of the Endangered Green Peafowl Pavo muticus in western Thailand: implications for conservation
- 396 guidance. Bird Conservation International 27: 1-17

397

398 Sukumal, N., Dowell, S.D., and Savini, T. in press. Modelling occurrence probability of the

399	endangered Green Peafowl in mainland Southeast Asia: applications for landscape conservation and
400	management. Oryx.
401	
402	Suwanrat, S., Ngoprasert, D., Sutherland, C., Suwanwaree, P., and Savini, T. (2015) Estimating
403	density of secretive terrestrial birds (Siamese fireback) in pristine and degraded forest using camera
404	traps and distance sampling. Global Ecology and Conservation 3: 596-606.
405	
406	Tilman, D. (1989) Ecological experimentation: strengths and conceptual problems. In: Likens, G.E.
407	(Ed.), Long-term Studies in Ecology. Springer, New York: 136–157.
408	
409	William, E.M., Philip, L.B., Brain, R.H., Leonie, C.M., and Johm, R.S. (2002) Population viability
410	analysis in endangered species recovery plans: past use and future improvements. Ecological
411	Application 12: 708–712.
412	
413	Zhang, Y., and Zheng, G. (2007) A population viability analysis (PVA) for Cabot's Tragopan
414	(Tragopan caboti) in Wuyanling, south-east China. Bird Conservation International 17: 151-161.
415	
416	

418	Figures
419	Figure 1. Predicted population size from 1992 to 2091 (blue line, year 1-100) of Green Peafowl in
420	HuaiKhaKhaeng Wildlife Sanctuary from a simulation model of population growth. Red shading
421	represents 95% credible intervals (CI).
422	
423	Figure 2. Predicted population size from 1998 to 2097 (blue line, year 1-100) of Green Peafowl in
424	YokDon National Park from a simulation model of population growth. Red shading represents 95%
425	credible intervals (CI).
426	
427	Figure 3. Predicted population size from 1992 to 2091 (blue line, year 1-100) of Green Peafowl in
428	HuaiKhaKhaeng Wildlife Sanctuary, by inputting 256 calling males of historical datasets at
429	population abundance in 1992 and inputting the mean number of $250 \pm 124$ (SD) calling males at
430	population abundance in 2015. Red shading represents 95% credible intervals (CI).
431	
432	Figure 4. Predicted population size from 1998 to 2097 (blue line, year 1-100) of Green Peafowl in
433	YokDon National Park, by inputting 475 calling males of historical datasets at population
434	abundance in 1998 and inputting the mean number of $450 \pm 145$ calling males at population
435	abundance in 2013. Red shading represents 95% credible intervals (CI).
436	
437	
438	
439	Figure 1



Years



443

444

445 Figure 3

